

1 **Genetic differentiation and diversity of the Bolivian endemic titi**
2 **monkeys, *Plecturocebus modestus* and *Plecturocebus olallae***

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22 **Abstract**

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24 The genetic variability of New World primates is still poorly documented. We present the
25 first genetic study on two threatened endemic titi monkey species in northern Bolivia
26 (*Plecturocebus modestus* and *Plecturocebus olallae*) using six microsatellite markers to
27 investigate genetic structure and variability of 54 individuals from two wild populations.
28 A low level of genetic diversity was found (34 alleles in the total sampled population).
29 Locus 1118 presented the greatest number of alleles. The mean number of alleles per
30 locus in the total population was 5.6 and the average heterozygosity was 0.38 (range:
31 0.12–0.88). The F_{IS} value for the total population using all microsatellite loci shows a
32 statistically significant heterozygote deficit. The inbreeding coefficients (F_{IS}) were
33 positive and significantly different from zero (0.064 for *P. olallae* and 0.213 for *P.*
34 *modestus*). The genetic differentiation between populations (F_{ST}) was moderate with a
35 pair-wise F_{ST} estimate of 0.14. Population structure analyses assigned the two populations
36 to two differentiated clusters ($K=2$). These results suggest that these two species with
37 very close distributional ranges arose from a single population, and that they remain in a
38 process of genetic differentiation and speciation. This study further underlines the urgent
39 need for conservation actions for both endemic primate species.

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41 **Key words:** Olalla's titi monkey; Beni titi monkey; microsatellites; genetic diversity;
42 conservation

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46 **1. Introduction**

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48 The study of genetic variation and population structure is critical to better understand
49 how species evolve, adapt and co-exist, as well as to propose rational conservation and
50 management strategies (Eguiarte, 1990). Molecular markers provide an estimate of
51 genetic diversity, with multiple alleles being useful to detect heterozygosity and
52 polymorphism levels in natural populations, determine inter and intra-population
53 relations, characterize genetic population structure, and analyse genetic distances for
54 evolutionary studies (Avice, 2000). As such, genetic studies are increasingly relevant to
55 taxonomy, biodiversity conservation and environmental monitoring programs to ensure
56 adequate management of species and populations (Groom, 2006).

57 The Olalla brothers' titi monkey (*Plecturocebus olallae*) and the Beni titi monkey
58 (*Plecturocebus modestus*) are two endemic primate species in Bolivia, which were
59 described by Lonnberg in 1939. Although their type localities are less than 65 km apart
60 (Figure 1), the respective specimens were consistently considered as belonging to two
61 different species in a series of taxonomic assessments for the entire Callicebinae group
62 (Hershkovitz, 1990; Kobayashi, 1995; Van Roosmalen *et al.*, 2002; Wallace *et al.*, 2013).
63 Studies on the distribution of both species have confirmed the presence of two similar,
64 but phenotypically distinct forms that correspond to the original descriptions with
65 distributional ranges quite close to each other (Figure 1; Felton *et al.*, 2006; Martínez &
66 Wallace, 2007, 2010, 2013, Martínez *et al.*, 2013). The two Bolivian titi monkey species
67 were considered Endangered by the IUCN (Veiga *et al.*, 2008^a, 2008^b), however in the
68 most recent assessment *P. olallae* was recognized as Critically Endangered (Martinez &
69 Wallace, 2016). These categorizations stem from the extremely restricted distributions

70 (especially *P. olallae*), habitat fragmentation, as well as low population density estimates
71 (Felton *et al.*, 2006; Martínez & Wallace, 2007, 2010, 2013; López-Strauss & Wallace,
72 2015).

73 A new taxonomic arrangement for titi monkey species was recently published in which
74 the original genus *Callicebus* was split up in three genera (*Cheracebus*, *Plecturocebus*,
75 and *Callicebus*). This decision was based on a molecular assessment of species
76 relationships and supported by biogeographical information that together explained the
77 species radiation process of titi monkeys in the Neotropics (Byrne *et al.*, 2016; 2018). We
78 follow this latest taxonomy for the two Bolivian endemic titi monkey species. Despite
79 substantial advances in taxonomical knowledge, there remains a lack of information for
80 species of the “donacophilus” group of titi monkeys which includes *P. olallae* and *P.*
81 *modestus*. Here, using molecular microsatellite markers, we present the results of an
82 analysis of the inter and intra-species genetic variation of the two Bolivian endemic titi
83 monkey species *P. olallae* and *P. modestus*.

84 **2. Methods**

85 *Study area*

86 We conducted our study in two localities. La Asunta is a cattle ranch found 2 km
87 northwest of the upper Yacuma River and 22 km from the original type locality of *P.*
88 *olallae*. The *P. modestus* study site was the Aguaizal community, approximately 45 km
89 from the original type locality (Figure 1). Both localities were known as titi monkey sites
90 from previous distributional studies (Martinez & Wallace 2017). The two localities were
91 found in central portions of the known distributional ranges of the titi monkey species
92 and separated by around 45 km.

93 *Preliminary assessment of faecal samples as DNA source*

94 We were committed to a non-invasive method to obtain DNA samples especially given
95 the risks that failures in procedures of live capture and release of individuals could have
96 on these threatened species. We identified faecal samples as a potential DNA source. To
97 assess the quality of faecal DNA samples we conducted a PCR amplification control
98 quality comparing faecal and blood samples from an individual of the geographically
99 neighbouring species, *Plecturocebus donacophilus*, held at the Zoologico de Fauna
100 Sudamericana, Santa Cruz de la Sierra, Bolivia. Extraction of DNA from blood tissue
101 was conducted using conventional CTAB-chloroform protocols (Doyle & Doyle, 1987;
102 Woodward *et al.*, 1994), modified and adapted at the Institute of Molecular Biology and
103 Biotechnology laboratory. The PCR amplification of all 12 microsatellite loci exhibited a
104 100% allelic correspondence across faecal and blood samples, highlighting the potential
105 of faecal samples as a DNA source.

106 As an additional confirmation, we found similar results from a comparison between
107 faecal samples obtained from our fieldwork (see details below) against muscular and
108 hairs samples from collected specimens of *P. modesuts* and *P. olallae* (Martinez *et al.*,
109 2013), respectively.

110 *Faecal sample collection*

111 We collected faecal samples from multiple individuals of titi monkeys belonging to 8
112 groups in La Asunta and 9 groups in Aguaizal. Titi monkeys are territorial primates
113 which emit early morning territorial calls (Kinzey 1988, Bicca Marquez & Heymann,
114 2013). We approached titi monkey groups while they were emitting these calls to ensure

115 all members were concentrated in a relatively small area thereby avoiding erroneous
116 assignation of individuals to other groups. At each encounter with a group we collected
117 faecal samples from 1-3 individuals by waiting for animals to defecate and immediately
118 collecting samples. This usually happened a few seconds after they noticed our presence.
119 In order to avoid multiple samples from the same individual, care was taken to ensure
120 that multiple sampled individuals from the same group were clearly distinguishable using
121 age-sex classes. Faecal samples were stored at ambient temperature in plastic bottles with
122 silica gel until transported to the laboratory.

123 *DNA isolation and genotyping*

124 Extraction of DNA was conducted using QIAamp DNA Stool Mini Kit (50) commercial
125 kit according to manufacturer specifications from collected primate faecal samples: 23 for
126 *P. olallae* and 31 for *P. modestus*. Given the lack of specific microsatellite markers for *P.*
127 *olallae* and *P. modestus*, we used seven heterologous markers developed for *Lagotrrix*
128 *lagotricha* (1110, 1118, 1115, 311, 312, 157, 113) (Di Fiore, 2004) and five markers
129 developed for *Alouatta palliata* (Ap 6, Ap 20, Ap 40, Ap 68, Ap 74) (Ellsworth, 1998).
130 Polymerase chain reactions (PCR) was conducted under standard conditions, using 40 ng
131 DNA in 1X of PCR buffer, 1.5 mM MgCl₂, 0.3 mM dNTP, 0.5 – 0.8 mM of each primer,
132 0.2 µg/ml of BSA and 0.05 U/µl –of Promega™ GoTaq™ DNA Polymerase and
133 additional ddH₂O for a final volume of 30 µL. The reaction cycle started at 95° C for 5
134 min; followed by 36 cycles of 35 sec at 94° C for denaturation, 30 sec at 50° C for primer
135 annealing (Di Fiore *et al.*, 2004), and 25 sec at 72° C for extension; and a final step of 10
136 min at 72° C. Each amplification reaction was performed four times for each sample. The
137 amplified fragments were separated in 10% polyacrylamide gel (ratio 19:1) and further

138 stained with silver nitrate. Allele sizes were estimated using a 25bp DNA Step Ladder,
139 and genotypes were checked four times. Quality control in genotype assignment was
140 performed with two examiners. Discrepancies identified through a custom-made database
141 were double-checked and resolved.

142 *Data analyses*

143 Genetic variability was evaluated according to the mean number of alleles per locus (A),
144 the mean number of effective alleles per locus (n_e), and the observed (H_o) and expected
145 (H_e) heterozygosity under the Hardy–Weinberg equilibrium, F_{IS} and their respective
146 means. These parameters were calculated using the GENETIX v. 4.05 32 software
147 (Belkhir *et al.*, 2004). Deviations from the Hardy–Weinberg equilibrium were tested
148 using the Markov chain method (100 batches of 1,000 interactions with a 1,000-step
149 dememorization process), applied to calculate unbiased estimates of exact probabilities (P
150 values). These tests were run on GENEPOP v.3.1c software (Raymond & Rousset, 1995).
151 Microsatellite loci were removed according to the type of data analysis as long as they
152 didn't have high frequencies of null alleles and strong deviations from Hardy-Weinberg
153 Equilibrium in both species. Estimations of null alleles and PIC (*Polymorphism*
154 *Information Content*) were performed with CERVUS 3.0.3 software. For this work, any
155 marker was removed. We determined differences between populations with the F_{ST}
156 statistic following Weir and Cockerham (1984) and genetic distance following Nei
157 (1972). Estimates of gene flow (Nm) were generated with the F_{ST} statistic. Genetic
158 distances, and factorial analyses of correspondence were conducted with GENETIX v.
159 4.05.

160 Individuals were assigned to populations using STRUCTURE v.2.0 (Pritchard *et al.*,
161 2000b). This method allows inferences about genetic population structure using
162 genotypes from a group of loci, as well as determining the probability that an individual
163 belongs to a given population. We used an algorithm that infers that the populations were
164 mixed. The value for the K parameter used in the analysis varied from 1 to 3 and the
165 number of iterations for each K was 50,000, with 10 replicate runs for each K using
166 correlated allele frequencies and an admixture model. Graphic representation of the
167 indicators of posterior probability of K cluster ($\ln Pr(X/K)$ and $\Delta(K)$) were obtained
168 with the web-based STRUCTURE HARVESTER software v.0.6.94 (Earl & VonHoldt,
169 2012).

170 **3. Results**

171 *Genetic diversity*

172 Six of the 12 microsatellite loci analysed were polymorphic (1118, 311, Ap40, 1115, 312,
173 Ap74), one was monomorphic (Ap68) and five resulted in poor amplifications (Ap6,
174 Ap20, 1110, 157, 113). Subsequent statistical analyses for both species were limited to
175 the six polymorphic microsatellite loci.

176 A total of 34 different alleles were observed across the 54 individuals of *P. olallae* and *P.*
177 *modestus* analysed, with a mean number of alleles of 5.6. The locus that presented the
178 greatest polymorphism was 1118 with 12 alleles, while the lowest amount of alleles was
179 for locus AP74 with 2 alleles (Table 1). The highest heterozygosity was observed for
180 locus 1118 ($H_o = 0.90$), whereas the lowest occurred on marker 312 ($H_o = 0.09$). The
181 locus showing the highest expected heterozygosity was 1118 ($H_e = 0.86$), whereas locus
182 312 had the lowest value for this parameter ($H_e = 0.15$). Three out of the six loci analysed

183 (311, 312, Ap 74) were identified to be in Hardy–Weinberg (HW) disequilibrium in the
184 global population (see Table 1). We found a statistically significant heterozygote deficit
185 (according to inbreeding F_{IS} value) for all the loci analysed except 1118 and Ap40 which
186 showed heterozygote excess, although not at a significant level (Table 1). The population
187 analysis showed only one marker in HW disequilibrium in both populations, therefore the
188 information of all markers was considered in subsequent genetic differentiation analyses.
189 The frequency of null alleles estimated ranged from 0 to 0.3849 while the polymorphic
190 information content (PIC) ranged from 0.15 to 0.84 (Table 1).

191 Both species showed the greatest number of alleles in the locus 1118 microsatellite
192 marker (Table 1). The mean number of alleles per locus was higher in *P. modestus* than
193 in *P. olallae* (Table 2). Seven exclusive or private alleles were identified for *P. olallae*
194 and nine for *P. modestus*, with different frequencies in the main markers (Table 2). The
195 observed and expected heterozygosity by species showed the average H_o ranged from
196 0.3146 (*P. olallae*) to 0.3656 (*P. modestus*), whereas the average H_e varied between
197 0.3283 (*P. ollalae*) and 0.4555 (*P. modestus*) (Table 2). In both species the heterozygote
198 deficit was significant (Table 2).

199 We found a low and nonsignificant heterozygote deficit (F_{IS}) for *P. olallae*, but it was
200 higher and statistically significant for *P. modestus*, indicating that the populations
201 assessed were not in a panmictic status (Table 2). No significant recent population
202 bottleneck was identified for either species.

203 *Genetic differentiation*

204 We found a low but significant genetic population structuring between the samples from
205 *P. olallae* and *P. modestus* ($F_{ST}=0.137$, $p<0.01$). Based on this result, the gene flow

206 between *P. modestus* and *P. olallae* indicates a theoretic migration of two individuals
207 between the two titi monkey species per generation ($Nm=1.57$).

208 According to the factorial analysis of correspondence, the first component explained
209 100% of the total variation and separated *P. olallae* and *P. modestus* individuals, showing
210 a clear tendency for separation of both species, forming two groups in each inertia axis,
211 with some overlapping individuals (Figure 2).

212 The results of the Bayesian analysis to assign individuals to groups indicated a clear
213 structure between the two titi monkey species. The highest likelihood was obtained when
214 K was set to two, with the distribution of $L(K)$ showing a clear peak for $K = 2$. Using the
215 method of correction of Evanno, the maximal K occurred at $K = 2$. Based on this and the
216 biological significance of the results, $K = 2$ was chosen as the final estimated number of
217 populations. The graphical representation of the clustering outcomes for $K = 2$ is shown
218 in Figure 3 with a few *P. modestus* individuals showing a high genetic relation with *P.*
219 *olallae*. According to these findings, the proportion of individuals correctly assigned to
220 each cluster was 90.4% for *P. olallae* and 79.2% for *P. modestus*.

221 **4. Discussion**

222 *Genetic variability*

223 This is the first study on the genetic variability, inter-specific genetic differentiation and
224 relationships between the Bolivian *P. olallae* and *P. modestus* whose restricted
225 distributional ranges are proximal (Martinez & Wallace, 2010, Wallace et al., 2013). We
226 obtained a larger number of alleles compared to previous titi monkey studies (Di Fiore,
227 2004), probably because of the larger number of sampled individuals ($n=54$). Di Fiore
228 (2004) reports failure to amplify the locus 312 for titi monkeys, but we amplified this

229 locus for both Bolivian *Plecturocebus* endemics showing better sequence compatibility.
230 For the primers which had not previously been tested for titi monkeys, we obtained few
231 amplified alleles for two of the loci (1 for Ap 68, 2 for Ap 74), and 5 alleles for Ap 40.

232 Our analysis based on the six DNA microsatellite loci revealed low allelic variability for
233 the two Bolivian titi monkeys (5.6 alleles for the entire population, 4.2 for *P. olallae* and
234 4.5 for *P. modestus*). The results described herein reveal low levels of genetic variability
235 comparable with previous studies performed on other species of monkey populations
236 using similar markers (Ellsworth, 1998, Di Fiore & Fleischer, 2004, Ruíz, 2007). This
237 may reflect the fact that the primers used are not specific to these species. Nevertheless,
238 the lack of correlation observed between the HWE deviation and the occurrence of null
239 alleles, suggest the existence of population subdivision (Wahlund effect), probably due to
240 a reduced effective population size within the studied species.

241 The differences in the number of identified private alleles (7 for *P. olallae* and 9 for *P.*
242 *modestus*) were especially useful for genetic differentiation between species, as their
243 presence suggests independent evolution. The total studied population had a mean
244 heterozygosity of 0.3709, while *P. olallae* presented a mean heterozygosity observed of
245 0.3146 and *P. modestus* 0.3656, values that are similar to the 0.33 mean reported for
246 *Plecturocebus moloch* (Alcantarino et al., 2009). Heterozygosity values of 0.5 or more
247 were reported for other threatened Neotropical primates such as *Alouatta pigra* (0.5;
248 García et al., 2005), *Lagothrix lagotricha* (0.8; Di Fiore & Fleischer, 2004). Therefore,
249 the values for the Bolivian titi monkeys suggest that their genetic variability might be
250 under ecological pressures.

251 Both *P. modestus* and *P. olallae* showed consanguinity, relatively low and not significant
252 for *P. olallae* ($F_{IS} = 0.064$), but higher and statistically significant for *P. modestus* ($F_{IS} =$
253 0.231), indicating that factors such as consanguinity, isolation, genetic drift or selection
254 may be acting on this species. Moreover, these results suggest a biased mating in the *P.*
255 *modestus* population between individuals with similar alleles. Population health can
256 decline due to consanguinity occurring in a few generations, a scenario that is an
257 especially serious conservation problem in small populations (Gilpin & Soule, 1986). The
258 consanguinity levels found may be a direct consequence of forest fragmentation present
259 across the entire distribution range of both endemic titi monkey species, with *P. modestus*
260 occupying a larger area, but in a drier forest, than *P. olallae* (Felton et al., 2006; Martinez
261 & Wallace, 2007, 2010, 2013). Drier habitat may be more sensitive to forest
262 fragmentation processes (Onderdonk & Chapman, 2000) promoting the spatial isolation
263 of *P. modestus* groups, thereby affecting their genetic variability.

264 Moreover, our sampling site for *P. modestus* is not as fragmented as the overall
265 distribution range of this species (Martinez & Wallace, 2010, Wallace et al., 2013) and
266 higher consanguinity levels could be expected for this species across its range. This
267 finding highlights the importance of population genetics information in the design of
268 conservation efforts, as combined with distribution and demographic information it
269 provides more details of a species conservation status.

270 Given the lack of significant evidence for recent bottlenecks in the studied populations,
271 the significant F_{IS} values identified in both studied populations may be the result of high
272 levels of inbreeding related to inbreeding depression and a reduction in the average
273 phenotypic values.

274 *Population structure*

275 The locus 1118 presented the largest number of alleles and the largest number of private
276 alleles for both species, with frequencies indicating a 68.75% possibility of finding one of
277 these alleles in *P. modestus*. The locus 1118 is therefore a good diagnostic marker for *P.*
278 *modestus*. The locus 311 presented 2 private alleles for *P. olallae*, with a 23.62%
279 possibility of finding them in individuals of *P. olallae*. The loci 1115 and 312 also had
280 low frequency private alleles for *P. modestus* (Table 3).

281 According to the genetic population structure that we found between the two species
282 ($F_{ST}=0.137$), 86% of the variation in allele frequencies is found within populations,
283 whereas 14% of the variance is attributable to differences between the populations. This
284 level of genetic differentiation suggests *P. olallae* and *P. modestus* could be recognized
285 as either subspecies or species (Holsinger et al., 2009). Additionally, although our results
286 show that a theoretical gene flow might exist between the two species and that divergence
287 due to genetic drift is low, the two populations do not behave as a single population,
288 which is a parameter that suggest a recent process of speciation (Chikhi et al., 2004).

289 The factor analysis of correspondence showed noticeable clusters for both endemic
290 *Plecturocebus* species. These results suggest that only a few individuals tended to
291 separate out from their population and lean towards the opposing species. The fact that
292 sampling localities are separated by around 45 km makes individual movements
293 unfeasible based on the knowledge about distribution, territoriality, and dispersion
294 patterns of the study species (Martinez & Wallace, 2010). Rather, this analysis supports
295 the hypothesis of a recent speciation event between the two titi monkey populations.

296 Genetic structure supports the previous statement, showing a clear distribution of the
297 population into two clusters when assigning two populations ($K=2$), with a reduced
298 number of shared genes, and a small number of migrants. When K was set as one ($K=1$),
299 thereby considering both species as a larger single population, individuals were assigned
300 in the same manner, forming two separate clusters with some individuals sharing genetic
301 characters. In addition, the alpha index value of 0.1213 indicates moderate genetic
302 population structuring, forming two separate populations, although certain individuals
303 still share alleles.

304 The patterns of genetic differentiation found in our study indicate that *P. modestus* and *P.*
305 *olallae* show genetic differences through characteristic private alleles, a distinguishable
306 distribution in a factorial analysis of correspondence and moderate but significant genetic
307 population structuring for the population. Nevertheless, we also found a number of
308 theoretic migrants (2) that still contribute to gene flow between the two species. Taken
309 together this suggests that these two species with very close distributional ranges arise
310 from a single population and that they are probably still in a process of genetic
311 differentiation and speciation.

312 The distribution range of *P. olallae* is found in the middle of the *P. modestus* range with
313 narrow but effective landscape scale barrier, in the form of seasonally flooded savannah,
314 impeding migration of individuals between populations (Martinez & Wallace, 2007;
315 2010, Wallace et al., 2013). This barrier might be relatively recent in evolutionary time-
316 scale, isolating a small population of titi monkeys, that later gave rise to *P. olallae*, in the
317 midst of a larger ancestral population that today we recognize as *P. modestus*. Genetic
318 parameters show these populations of *Plecturocebus* as clearly different at a subspecies

319 level, but the evidence also suggests an apparent recent speciation process, which when
320 combined with already identified morphological and phenotypic distinctions, suggests
321 that *P. olallae* and *P. modestus* are two very close species still undergoing genetic
322 divergence.

323 We found a higher level of consanguinity for *P. modestus*, which may signify a higher
324 extinction risk than for *P. olallae*. Given that *P. modestus* has a larger distributional range
325 and population size than *P. olallae* (Martinez & Wallace, 2007, 2010; Lopez-Strauss &
326 Wallace, 2015), these observations were unexpected. However, larger populations do not
327 necessarily have higher genetic variability and less consanguinity (Lacy, 1987). The
328 forests are fragmented across the range of both species, and our results may simply reflect
329 the population dynamics on each specific study site, although the *P. modestus* locality
330 had larger and more continuous forest coverage than the *P. olallae* site. From a genetic
331 perspective, our preliminary results suggest that *P. modestus* may be in a critical situation
332 as compared to *P. olallae*. Future studies should consider including mitochondrial
333 markers and expanding sampling sites across the distribution ranges of each species.

334 The IUCN previously classified both *P. olallae* and *P. modestus* as Endangered due to
335 small distribution ranges, habitat fragmentation and habitat loss risks, as well as
336 population decline (Martinez & Wallace, 2007, 2010, 2013; Veiga et al., 2008a, 2008b;
337 Wallace et al., 2013; López-Strauss & Wallace, 2015). More recently, *P. olallae* was re-
338 categorized as Critically Endangered due to an extremely limited distribution (Martinez
339 & Wallace, 2016). From a genetic perspective, *P. modestus* faces a greater risk of loss in
340 genetic quality than *P. olallae*, which should be considered in further conservation
341 actions.

342 **Acknowledgements**

343 We thank the Wildlife Conservation Society, Gordon and Betty Moore Foundation,
344 Primate Conservation Inc., Margot Marsh Biodiversity Foundation, BP Conservation
345 Leadership Program, and Conservation International Primate Action Fund for financial
346 support for conservation research efforts on the Bolivian *Plecturocebus* endemics. We
347 thank our field assistant Hector Cáceres, and also Lesly Lopez who helped in data
348 collection. We thank the National Directorate for the Protection of Biodiversity for help
349 in acquiring necessary research permits, as well as the collaboration of the Municipalities
350 of Reyes, San Borja, and Santa Rosa del Yacuma, and the Nogales cattle ranches for
351 access to the study sites. We acknowledge the support of the Institute of Biology
352 Molecular and Biotechnology (Universidad Mayor de San Andres) in La Paz.

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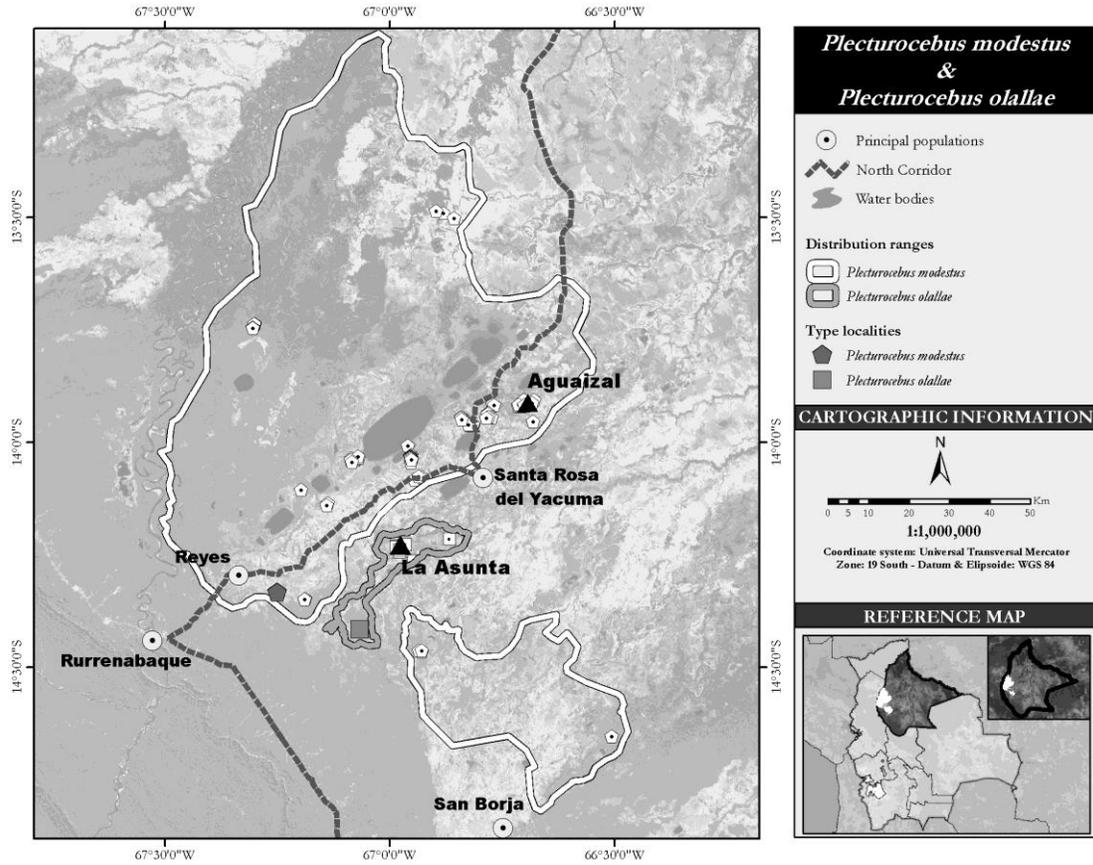
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464 **Figures**

465 **Fig. 1** Map showing known distribution of *Plecturocebus modestus* and *Plecturocebus*
466 *olallae*, original type localities and study samples localities.

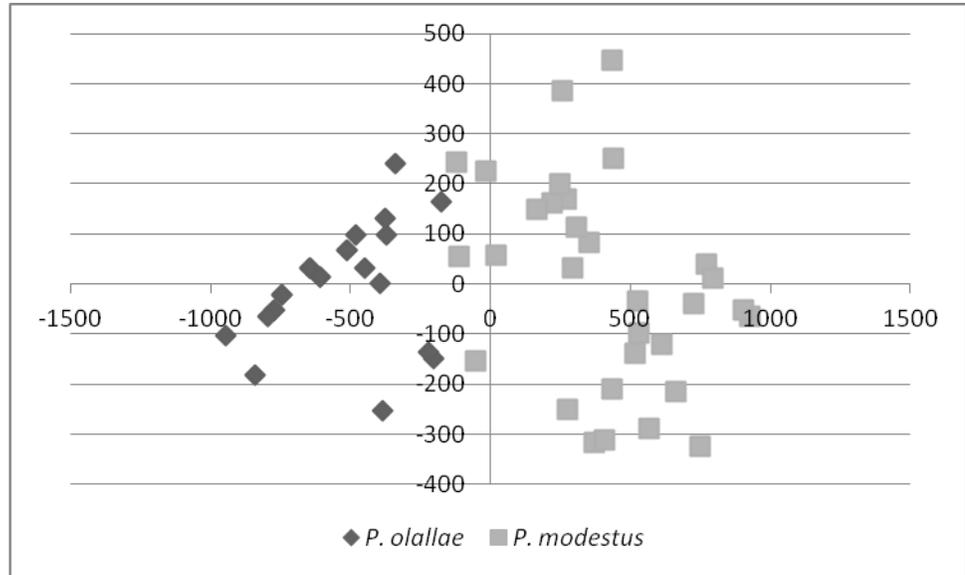
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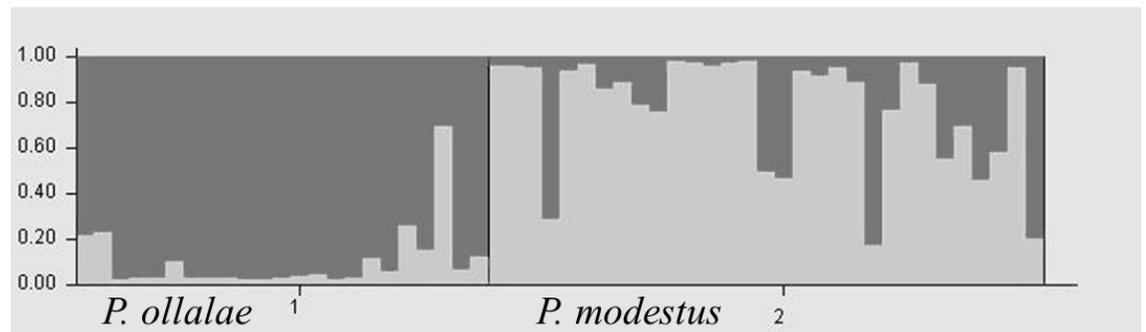
Fig.2 Positioning of individuals within the Cartesian plane constructed from factorial analysis of correspondence of *Plecturocebus olallae* (dark gray) and *Plecturocebus modestus* (light grey)



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479 **Fig.3.** Graphical representation of the clustering outcomes suggested by the Bayesian
480 analysis performed to assess the structure of the studied populations at $K = 2$. Each color
481 represents one cluster, and the length of the colored segment shows the individual's
482 estimated proportion of membership in that cluster. Black lines separate the individuals
483 of the different populations.



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487 **Table 1.** Genetic diversity parameters estimated for the 6 microsatellite markers analyzed
488 in the two *Plecturocebus* species considered in this study

Marker	A ^a	Ho ^b	He ^c	HWE ^d	Fis ^e	F (null) ^f	PIC ^g
1118	12	0.9057	0.8610	ns	-0.0423 ^{ns}	0.0335	0.847
311	7	0.4151	0.7577	***	0.4597***	0.2220	0.717
Ap 40	5	0.2778	0.2476	ns	-0.1126 ^{ns}	0.0000	0.231
1115	4	0.1852	0.2368	ns	0.2268*	0.0000	0.228
312	4	0.0926	0.1562	*	0.4150**	0.1089	0.150
Ap 74	2	0.1852	0.3656	***	0.5004**	0.3849	0.299
Total	34						

489 ^a A: number of alleles per locus.

490 ^b Ho: average observed heterozygosity.

491 ^c He: average expected heterozygosity.

492 ^d HWE: significant deviation from the Hardy–Weinberg equilibrium (P<0.001).

493 ^e Fis: coefficient of inbreeding.

494 ^f F (null): frequency of null alleles estimated for each locus.

495 ^g PIC: polymorphic information content.

496 *P<0.05.

497 ** P<0.01.

498 *** P<0.001.

499 ns: no significant

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504 **Table 2.** Genetic variability parameters estimated for two species of *Plecturocebus* based
505 on the analysis of 6 microsatellite markers

Species	n ^a	Ho ^b	He ^c	HEW ^d	MNA ^e	pA ^f	Fis ^g
<i>P. ollalae</i>	23	0.3146	0.3283	*	4,2	7	0.06484
<i>P. modestus</i>	31	0.3656	0.4555	**	4,5	9	0.21311** *

506 ^an: sample size.

507 ^bHo: average observed heterozygosity.

508 ^cHe: average expected heterozygosity.

509 ^dHWE: significant deviation from the Hardy–Weinberg equilibrium (P<0.001).

510 ^eMNA: mean number of alleles.

511 ^fpA: number of private alleles.

512 ^g Fis: estimates and significance of the deviation from Hardy-Weinberg equilibrium per
513 population across the 6 loci.

514 *P<0.05.

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Table 3 Frequency of private alleles per locus for each species

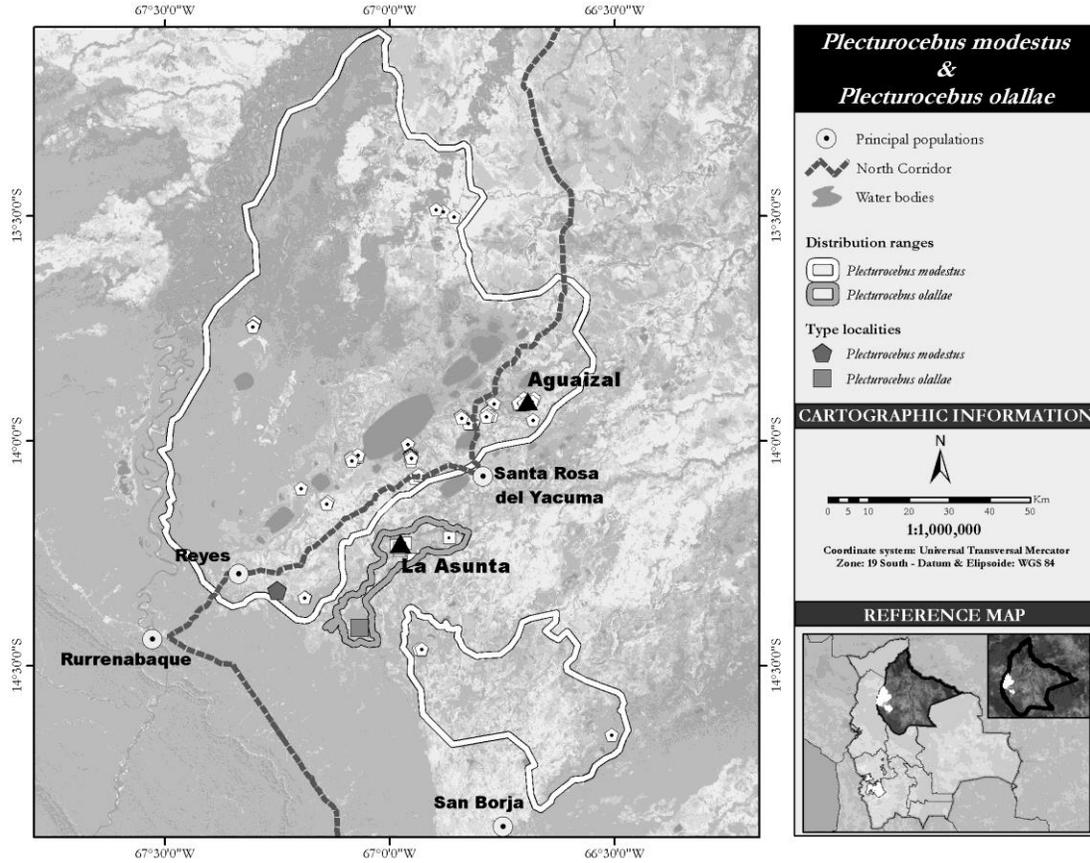
Species	Locus	Allele	Frequency
<i>P. modestus</i>	1118	162	0.3875
		174	0.1375
		178	0.1625
	1115	226	0.0875
	312	162	0.0750
<i>P. olallae</i>	1118	164	0.0556
		186	0.0833
	311	176	0.1806
		180	0.0556

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1 **Figures**

2 **Fig. 1** Map showing known distribution of *Plecturocebus modestus* and *Plecturocebus*
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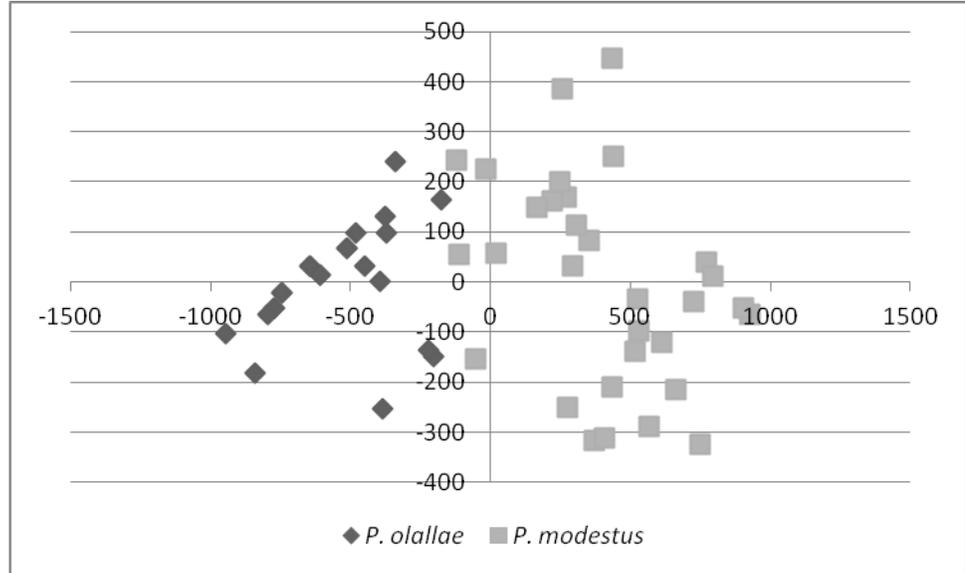


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7 **Fig.2** Positioning of individuals within the Cartesian plane constructed from
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9 *Plecturocebus modestus* (light grey)

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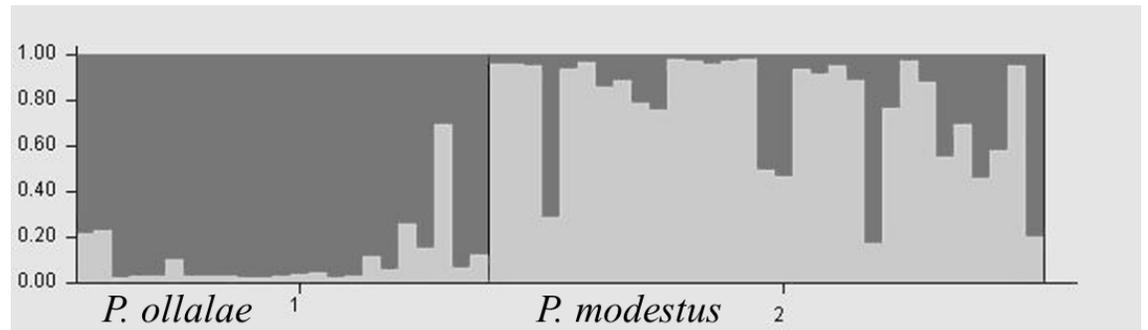
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16 **Fig.3.** Graphical representation of the clustering outcomes suggested by the Bayesian
17 analysis performed to assess the structure of the studied populations at $K = 2$. Each color
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21

1 **Table 1.** Genetic diversity parameters estimated for the 6 microsatellite markers analyzed
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9 ^g PIC: polymorphic information content.

10 *P<0.05.

11 ** P<0.01.

12 *** P<0.001.

13 ns: no significant

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18 **Table 2.** Genetic variability parameters estimated for two species of *Plecturocebus* based
19 on the analysis of 6 microsatellite markers

Species	n ^a	Ho ^b	He ^c	HEW ^d	MNA ^e	pA ^f	Fis ^g
<i>P. ollalae</i>	23	0.3146	0.3283	*	4,2	7	0.06484
<i>P. modestus</i>	31	0.3656	0.4555	**	4,5	9	0.21311** *

20 ^an: sample size.

21 ^bHo: average observed heterozygosity.

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		186	0.0833
	311	176	0.1806
		180	0.0556

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